

1 **Proposing a socioecological framework for successful grassland restoration in**
2 **Germany – an overview and insights from the *Grassworks* project**

3

4 Vicky M. Temperton ^{*1}, Ioana A. Pătru-Dușe¹, Alina Twerski¹, Philipp Laeseke², Regina Neudert³,
5 Miguel Cebrian-Piqueras¹, Manuel Pacheco Romero¹, Markus Bauer⁴, Volker Beckmann³, Jörn
6 Fischer¹, Konrad Gray⁵, Werner Härdtle¹, Johannes Kollmann⁴, Lukas Kuhn¹, Christin Juno
7 Laschke¹, Jacqueline Loos⁶, Lotte Lutz¹, Felix May⁷, Michaela Meyer⁸, Berta Martin-Lopez¹,
8 Maraja Riechers⁹, Moritz Ptacek⁴, Desirée Seifert⁸, Annika Schmidt⁵, Line Sturm⁵, Jan Thiele²,
9 Sabine Tischew⁵, Liselotte Unseld⁸, Terese Venus¹⁰, Miriam Wiesmeier⁴, Anita Kirmer⁵

10

11 Affiliations:

12

13 1 Faculty of Sustainability, Leuphana University Lueneburg, Germany

14 2 Thünen Institute for Biodiversity, Braunschweig, Germany

15 3 Faculty of Law and Economics, University of Greifswald, Germany

16 4 Department of Life Science Systems, TUM School of Life Sciences, Freising, Germany

17 5 Hochschule Anhalt University of Applied Sciences, Bernburg, Germany

18 6 Department of Botany and Biodiversity Research, Faculty of Life Sciences, University of Vienna, Austria

19 7 Institute of Biology, Free Universität Berlin, Germany

20 8 German Association for Landcare (DVL), Ansbach, Germany

21 9 Thünen Institute of Baltic Sea Fisheries, Rostock, Germany

22 10 Faculty of Business, Economics and Information Systems, University of Passau, Germany

23 *Corresponding Author: Vicky M. Temperton (vicky.temperton@leuphana.de)

24 **Abstract**

25 Bending the biodiversity curve and delivering on biodiversity promises from international
26 agreements and laws, including Kunming-Montreal and the EU Restoration Law, requires
27 upscaling ecological restoration from smaller to larger spatial and temporal dimensions and
28 across different spheres of society. Achieving this depends on a strong scientific evidence base
29 and synthesis of effective practices from both ecological and social perspectives.

30 The *Grassworks* project investigates the factors driving success in grassland restoration in
31 Germany, addressing ecological, socio-economic, and socialecological dimensions. We address
32 this by conducting a post-hoc assessment of previously restored sites, comparing them to both
33 positive and negative reference sites across three regions along a north-south gradient in
34 Germany. In the post-hoc assessment, we employed a stratified design to evaluate the effects
35 of restoration methods, previous land use, current management, governance, finance, and time
36 since restoration intervention. We assessed vegetation, butterflies, wild bees, soil
37 characteristics, and economic performance, while controlling for surrounding landscape
38 configuration. Additionally, we examined key socialecological dimensions, including stakeholder
39 values, knowledge exchange, and decision-making processes within established networks. This
40 was complemented by a Real-World Laboratory approach, integrating ex-ante and ex-post
41 assessments, demonstration sites, and live restoration activities co-created with local
42 stakeholders.

43 This publication provides an overview and reflection, drawing on insights from the *Grassworks*
44 project in Germany, to inform, guide and support the development of future large-scale
45 socialecological restoration efforts worldwide.

46

47

48 **Keywords:** governance, grassland, multifunctionality, open ecosystems, production economics,
49 restoration success, socialecological approach, stakeholder engagement, transdisciplinarity

50

51

52 **Implications for Practice**

- 53 • Legal framings such as the EU Restoration Law or the Kunming-Montreal Biodiversity
54 Agreement now require a strong evidence base for scaling up restoration initiatives on
55 the ground. Applying a socialecological lens to what constitutes success in restoration as
56 well as integrating findings across regions promise to contribute significantly to
57 providing such a science and practice driven evidence base. The socialecological and
58 landscape-level approach used in *Grassworks* can be replicated in other large-scale
59 restoration research to connect scientific findings more strongly with policy and
60 practice.
- 61 • Such an integrative approach requires frequent communication and a common language
62 as well as trust between scientists from disparate fields of enquiry (across natural and
63 social sciences). This required a high level of openness and as well as time. The level of
64 communication and amount of time required is rewarded with more generalisable,
65 broad, robust and integrative outcomes, however.

66

67 **Introduction: setting the scene**

68 Species-rich grasslands worldwide are under threat of degradation despite grasslands providing
69 a wide array of different ecosystem functions and therefore also benefits to people (Bengtsson
70 et al. 2019, Bardgett et al. 2022), and covering more than a third of terrestrial land globally
71 (Squires et al. 2018). As the UN Decade on Ecosystem Restoration unfurls, scaling up
72 restoration activities (Shackelford et al. 2013; Perring et al. 2018) represents one of the main
73 opportunities and challenges. Developing a strong evidence base for best practice in restoration
74 will form a key component of scaling up.

75 Restoration activities on grasslands have, until recently, mainly focused on achieving ecological
76 targets such as increasing plant diversity (Tischew et al 2008; Helm et al. 2015; Hess et al. 2020;
77 Shackelford et al. 2021a) or more recently improving biotic interactions such as plant–pollinator
78 networks (Montoya et al. 2012; Traveset et al. 2023). Socio-political, cultural or economic
79 factors have so far received limited attention in the assessment of restoration success (but see
80 Wortley et al. 2013; Fernández-Manjarrés et al. 2018; Elias et al. 2021; Tedesco et al. 2023),
81 despite evidence that ecologically successful projects are often influenced by social framings,
82 such as the acceptance of restoration outcomes (Pfadenhauer 2001), cost considerations
83 (Waldén & Lindborg 2018), governance, implementation and protection (Canessa et al. 2023).
84 Although research on payments for ecosystem services in grasslands has gained traction, and
85 result-based payments represent a method of evaluation of restoration success and economic
86 incentives (Huber & Finger 2020), integrative approaches that combine economic with
87 ecological and socialecological factors remain rare. Examples of social framings influencing
88 restoration success, include ecologically valuable wildflower meadows being threatened with
89 destruction in urban areas where the city authorities consider selling the land for building
90 (personal communication), or tree planting initiatives being ecologically sound but failing due to
91 the planting activities occurring at the wrong time of the year (Messier et al. 2014). At the same
92 time, these projects or initiatives significantly affect the lives of individuals, influencing the
93 development of local economies, the configuration of governance structures, and the cultural
94 connections to restored landscapes. The extent to which stakeholders – such as farmers,
95 landowners, conservationists, local community members, policymakers, and restoration

96 practitioners – are included in (i) restoration actions, decision making processes, levels of
97 participation and power dynamics, (ii) the extent to which the connection to nature motivates
98 practitioners and stakeholders to restore grasslands, and (iii) the socio-economic and policy framing of
99 activities can and likely does significantly influence the success of restoration. These aspects of
100 restoration have been understudied (see Broeckhoven & Cliquet 2015; Martin 2017; Fischer et
101 al. 2021; Buckingham et al. 2021), while the awareness of the importance of such social factors
102 increases. Studies that include equal focus on the ecological and the social, as well as the
103 interface between the two, are so far practically non-existent (hence the appeal in Fischer et al.
104 2021, to perform such integrative research in the UN Decade on Ecosystem Restoration; see
105 also Tedesco et al. 2023).

106 In this inter- and transdisciplinary project, *Grassworks*, we aim to holistically fill this research
107 gap by investigating under what conditions grassland restoration is successful, explicitly
108 including ecological, socialecological as well as socio-economic variables. In large parts of
109 Europe, species-rich grasslands are among the most threatened habitat types, with <10% of
110 these grasslands that are protected under EU law being in a favourable condition, and 75%
111 showing negative trends (Wesche et al. 2012; see also Dengler & Tischew 2018). Studies in
112 Germany have highlighted not only the extent of plant species loss, but also the specific types
113 of species being lost, with the majority being those adapted to open ecosystems, particularly
114 grasslands (Jandt et al. 2022; Staude et al. 2023). Therefore, restoring species-rich meadows
115 and pastures, as emphasized in the recently ratified EU Restoration Law
116 (https://environment.ec.europa.eu/topics/nature-and-biodiversity/nature-restoration-law_en),
117 is highly likely to deliver a substantial societal return on investment (De Groot et al. 2013;
118 Shipley et al. 2020).

119 In the *Grassworks* Project, we address the central research question: 'What leads to success in
120 grassland restoration in Germany?' by exploring ecological, socio-economic, and
121 socialecological dimensions. We propose a synthetic integration (*sensu* Fischer et al. 2021)
122 grounded in the insights and methodologies developed within the project. Through this
123 comprehensive approach, we aim to establish a robust evidence base to inform and guide
124 future socialecological restoration efforts (<https://grassworksprojekt.de/en/>). The project

125 integration focuses on three regions in Germany, thus providing a latitudinal gradient from
126 North to South. As a backdrop to our findings, we acknowledge the well-documented latitudinal
127 biodiversity gradient in European grasslands, with higher biodiversity observed in southern
128 regions (Dengler et al. 2014). In the *Grassworks* project, we consider this expected gradient as a
129 contextual baseline for exploring restoration success, while recognizing that regional
130 differences often outweigh these broader macroecological patterns (Stoher et al. 2013). We
131 explicitly selected a wide range of grassland habitats to ensure the outcomes are both more
132 generalizable and more easily transferable to other grassland systems.

133 This publication has two main goals: firstly, **it provides an overview of the approach and**
134 **methods used to assess ecological, socialecological and socio-economic restoration success**
135 across three different regions in Germany (North, Centre, South). Secondly, given the extensive
136 efforts to standardize and improve comparability of the measurements across restored sites
137 and regions, **we provide insights into the decision-making and reflection processes that**
138 **underlie the experimental design and approaches taken.** This includes the main rationale of
139 the research design, with the aim of allowing other researchers to replicate or refine and adapt
140 the design further, and fostering continuous improvement by advancing the evidence base for
141 restoration success. This highlights the critical importance of integrating social and economic
142 dimensions into restoration efforts and success, thereby recognizing the interdependence of
143 ecological, economic and social phenomena.

144 Such projects are directly relevant to policy and practice as they help identify intervention
145 points that target key socialecological transformations (see 'leverage points' in Abson et al.
146 2017) and guide future cost-effective strategies to maximize the likelihood of restoration
147 success.

148 **Design of the *Grassworks* project**

149 From a socialecological perspective, the most successful grassland restoration can be defined as
150 a process that considers the ecological components, social aspects and socio-economic facets,
151 as well as the improved benefits to people. Ecologically, we perceive grassland restoration as
152 successful when as many native grassland species as possible are established, leading to higher

153 alpha, beta and gamma diversity as well as improving vegetation structure and ecosystem
154 functions (in other words enabling “ecological complexity”; see Wortley et al. 2013).
155 Restoration projects increasingly also consider the forb-grass ratio (that is linked to multitrophic
156 interactions) as a key metric of restoration success (Bucharova et al. 2020; Nerlekar et al.
157 2024). Reflecting these principles, our project included the assessment of species diversity and
158 vegetation structure, including the forb–grass ratio, to evaluate restoration success more
159 comprehensively.

160 Given the acceleration of global change as well as the need to include a wider range of human-
161 related outcomes, restoring ecosystem functions and services is an emerging focus of
162 socialecological restoration (Funk et al. 2017; Carlucci et al. 2020). From a social perspective,
163 successful restoration also improves human–nature connections (that integrate diverse values,
164 practices and knowledge) and achieves a balance between natural processes and human needs,
165 combined with inclusive governance and effective economic incentives across temporal and
166 spatial scales (Fischer et al. 2021; Tedesco et al. 2023). Finally, these factors contribute to the
167 resilience of the system, so it is sustained for future generations (Lyons et al. 2023).

168 Restoration is very likely to deliver diverse benefits to the society, but it is inevitably connected
169 with costs for initial restoration and maintenance which are often borne by farmers and
170 landowners (Zerbe 2023). From a socio-economic perspective, restoration success translates at
171 the societal level into positive social net-benefits or high benefit-cost ratios, indicators which
172 require that all costs and benefits to all members of the society are measured in monetary
173 terms, and are hardly calculated (for an exception, see De Groot et al. 2013). Cost-effectiveness
174 measures might be used if benefits, like species richness, are not measurable in monetary
175 terms (Knight & Overbeck 2021). At the site or farm level, profitability, cost coverage,
176 employment and income, and long-term maintenance perspective are typical indicators of
177 successful restoration (Waldén & Lindbog 2018, Ben-Othmen & Ostapchuk 2023).

178
179 The *Grassworks* approach combines a post-hoc assessment of already restored sites with real
180 world laboratories (RWLs) in each of the three study regions. This follows recommendations in
181 Fischer et al. (2021) for a research agenda for socialecological restoration in the UN Decade on

182 Ecosystem Restoration. Our central hypothesis is that restoration success relates to the extent
183 to which both ecological complexity (encompassing biodiversity, vegetation structure,
184 ecosystem functions) and social engagement (stakeholder diversity, inclusion) are considered in
185 the restoration process. We hypothesise that the higher the ecological complexity and social
186 engagement are, the higher the restoration success will be (Figure 1). To maximise the potential
187 for restoration success, *Grassworks* is creating an integrative framework that can be used to
188 identify potential scenarios for how ecosystem multifunctionality can be enhanced through the
189 process of grassland restoration.

190

191 **Main Rationale of the Study Design: three regions as a natural landscape experiment and a**
192 **real-world lab for transformation and learning**

193 To assess the success of restoration projects, we used a landscape scale ‘natural experiment’
194 design with an assessment of ecological, economic and social attributes of restoration in 187
195 sites across three different regions in Germany (Fig. 2).

196 The Northern study region is economically (measured in GDP per capita) and ecologically in the
197 median range (related to plant species richness) compared to the whole of Germany. The
198 Central region combines good ecological quality with lower economic strength, and the
199 Southern region combines strong economic performance with above-average ecological quality
200 (Peisker 2023; [https://de.statista.com/statistik/daten/studie/73061/umfrage/bundeslaender-
201 im-vergleich-bruttoinlandsprodukt/](https://de.statista.com/statistik/daten/studie/73061/umfrage/bundeslaender-im-vergleich-bruttoinlandsprodukt/)). For both the Centre (Saxony-Anhalt) and South (Bavaria)
202 all *Grassworks* sites lay within one federal state, whereas for the North we sampled in four
203 different federal states, namely Lower Saxony, Schleswig Holstein, Hamburg and northern parts
204 of Saxony-Anhalt, each with partially varying agri-environment and impact mitigation schemes,
205 as well as differing economic conditions.

206

207 **Post-hoc assessment: study design and landscape experimental set up**

208 We developed a post-hoc assessment to provide a holistic analysis of factors that affect
209 grassland restoration success in Germany, collecting ecological, socialecological and economic
210 data in grasslands already restored by local stakeholders in the three regions (Figure 2). Data

211 was collected over two growing seasons, spanning 2022 and 2023. In addition to measuring
212 local site conditions, we used remote sensing data to assess the surrounding landscape around
213 the restored sites by delineating different land use types (grassland, arable land, forest,
214 settlements and others) and compiling plant species richness for each land use type within a
215 300 m radius of each restored site.

216 Overall, we sampled 121 restored grassland sites, as well as 33 negative and 33 positive
217 reference sites across Germany, giving a total of 187 sites (with around 40 restored sites plus
218 ten positive and ten negative sites per region; Table 2). We included dry, fresh and moist to wet
219 grassland types (excluding only grasslands on peat soils). Target vegetation types measured
220 (referenced in Chytrý et al. 2020) were semi-dry calcareous grassland (R1A), pastures (R21),
221 lowland hay meadows (R22), moist or wet eutrophic meadows (R35) and moist or wet
222 oligotrophic grasslands (R37). The grassland sites represent a wide gradient of different
223 conditions in terms of their ecological and socialecological characteristics. Our design was
224 chosen to increase transferability of results across Germany and to similar temperate
225 conditions. In line with many ecological restoration projects, we compared variables measured
226 in restored sites with positive reference (non-degraded) as well as negative (degraded)
227 reference sites (*sensu* Zedler 2007; Wortley et al. 2013) see Fig. 2, also Box 1 and 2 for a
228 reflection on the process of site selection). This approach allows a comparison of the variability
229 of factors that affect restoration success within and between three larger regions and should
230 significantly increase the predictive power of such studies for restoration measures (*sensu*
231 Brudvig 2017). The final randomized stratified design included the following main factors that
232 can influence restoration outcomes, i.e. restoration method, age since main restoration
233 intervention, previous land use and current management (grazing or mowing, or a combination
234 of both).

235

236 **Post-hoc assessment: ecological variables and field sampling**

237 Each site was sampled once per year using a space-for-time approach, to assess vegetation, soil
238 chemistry and texture. At each site, a 200 m transect (5 m wide) was set up and marked using

239 GPS coordinates, with an accuracy of 0.01 m (Figure 3). Each transect was separated into four
240 50-m sub-transects that were sampled for vegetation, soil and insects.

241 *Vegetation:* within each 50-m section, 4 m² (2 m x 2 m) vegetation plots were surveyed
242 positioned at points derived at random along the 50 m stretch, with the minimum distance of 5
243 m from the end of the transect), giving four plots per site and a total vegetation sampling area
244 of 16 m² (Figure 3). Within each vegetation plot we assessed species presence as well as cover
245 using a modified Braun-Blanquet scale (see Table S1 for details). Additional plant species were
246 recorded on the whole 1000 m²-transect. The maximum duration allocated for sampling
247 additional plant species was one hour. Additionally, vegetation height was measured four times
248 per year using a drop disc along the 200 m transect. To assess the percentage of area covered
249 by flowers, overhead photos of the vegetation were taken within the 4 m² quadrants during the
250 insect surveys and subsequently analyzed in the lab.

251
252 *Soil:* At each site in March or early April at each vegetation plot, we took soil samples (pooled
253 from six soil cores, 20 mm diameter) that were further pooled into one sample per site and
254 analysed for total soil organic carbon (SOC), total nitrogen content, pH and soil texture as well
255 as microbial biomass (carbon-based). Additionally, soil bulk density was measured at two
256 locations per site, to enable future assessment of carbon sequestration over time. The soil and
257 bulk density samples were taken at two depths, namely 0–10 and 10–30 cm since these depths
258 are commonly sampled across Germany and allow for national and international comparison.

259
260 *Insects:* Butterfly and wild bee sampling was done monthly, four times per site from May to
261 August along the 200 m transect (width: butterflies 5 m, wild bees 2 m) when weather
262 conditions were suitable (see Figure 3). Butterflies were counted using the Pollard walk method
263 (Pollard 1977), and wild bees were collected by sweep netting for 5 min within each 50-m sub-
264 transect section, resulting in a total of 20 min of observation per transect. Additional butterfly
265 and wild bee species were collected by conducting two further 5-min random walks across the
266 entire site. Butterflies were identified to species level in the field and wild bees were collected
267 and identified in the laboratory. This overall ecological and biophysical sampling constitutes an

268 elaborate range of variables assessed using standardised methods consistently applied across
269 all three regions.

270

271 **Post-hoc assessment: surrounding landscape, production economics and socialecological** 272 **dimensions**

273 In addition to the ecological variables measured *within* all the sites, we assessed the
274 surrounding landscape using land-cover datasets from authorities (Table S3), satellite images as
275 well as on-the-ground assessment of plant species richness within a 300 m radius around the
276 sites. These data allow us to also explore the relation between the surrounding landscape and
277 restoration outcomes, including assessing the possible role of available extensive grassland area
278 in the surrounding landscape, landscape diversity (the number and share of different land use
279 components) and landscape configuration. We created landscape data for a radius of 2 km
280 around each site.

281 Land-cover datasets were aggregated and checked for missing objects and errors using the
282 digital orthophoto with 20-cm resolution (DOP20). For example, we digitized missing landscape
283 features such as hedges and ditches with at least 2 m width (Table S3). In addition, we used the
284 crop type and mowing events raster layers by Blickensdörfer et al. (2022) and Schwieder et al.
285 (2022), respectively, to calculate area of extensive grassland in the surrounding and crop-type
286 diversity, as well as amount of available pollen and nectar and pesticide use (following Hellwig
287 et al. 2022). In total, we collected and produced data for 1916 km². All geographical data were
288 processed in R (version 4.1.2; R Core Team) with the packages: sf (Pebesma & Bivand 2023;
289 Pebesma 2018), terra (Hijmans 2024), osmdata (Padgham et al. 2017), and the tidyverse
290 (Wickham et al. 2019). Corrections were made in QGIS (version 3.28.0-Firenze).

291

292 All sites were incorporated into a production-economics assessment using online
293 questionnaires to gather data on initial restoration efforts and current management practices.
294 This assessment aimed for an in-depth cost-coverage and cost-effectiveness analysis of the
295 measures implemented throughout the initial restoration and management phases. The
296 questionnaire on initial restoration (implemented in the software Unipark, Tivian 2024)

297 included questions on the timing and methods of restoration, including soil preparation, seed
298 introduction, initial maintenance and financing. A questionnaire on current management
299 requested information on type of management, timing, utilisation of forage, maintenance
300 measures and financial support for the year 2022. Relevant stakeholders (farmers,
301 administrative and NGO personnel) were approached to fill out questionnaires online or on the
302 phone from January 2023 to March 2024. Furthermore, we developed a broader
303 sociaecological assessment based on stakeholder's perceptions of the type of restoration
304 performed, the restoration goals and success, and the effect of the sociaecological context.
305 Regarding the type of restoration, we explored key aspects such as the degree of stakeholder
306 involvement, the type of knowledge applied, or the approach and practices used. We also
307 explored the perceived level of priority as a restoration goal and the achieved success of
308 aspects related to plant and insect diversity, the degree of human-nature connections,
309 livelihood opportunities or social cohesion, among others. Within the sociaecological context,
310 we assessed if factors such as the climate conditions, land-use practices, stakeholder
311 engagement or people's values towards restoration were perceived as enhancers or inhibitors
312 of the restoration process. The perceptions were collected through an online questionnaire that
313 was sent to stakeholders with different roles and degrees of involvement in the restoration
314 process. We used multivariate analyses to identify archetypes of restored grasslands based on
315 restoration type, prioritization of goals, perceived success, and the influence of the
316 sociaecological context. These archetypes summarized stakeholder perceptions of the
317 restoration process and the balance between ecological and social success, providing a
318 complementary perspective to ecological field data and production economics analyses. In
319 addition to asking the stakeholders directly responsible for the implementation of restoration in
320 the post hoc sites, we assessed the public value of grassland restoration through surveys with
321 the public across Germany. Our framework highlighted the diverse values driving stakeholder
322 engagement, emphasizing their role in fostering inclusive and effective restoration efforts.

323 **Real-world lab as a transdisciplinary approach**

324 At the heart of *Grassworks'* socioecological research component are the real-world laboratories
325 (RWL). Working at the interface between science, practice and local communities, RWLs are
326 becoming more widely used for exploratory and transdisciplinary research approaches
327 (Schäpke et al. 2018; Bergmann et al. 2021). The main function of the RWLs is to act as newly
328 structured forms of cooperation and collaboration between scientific and social actors, as well
329 as open spaces for research, social learning and co-design, where new ideas can be thought
330 through in a transdisciplinary and experimental way, thereby initiating transformation
331 processes (Schäpke et al. 2024). As part of our project, the RWLs provided an experiential
332 environment where we engaged with various stakeholders in shared learning, co-creation and
333 reflection on their practices and future perspectives related to grasslands. It is important to
334 note that RWLs are inherently contextual and normative, collaboratively designed to identify
335 and address potential sustainability challenges (Wanner et al. 2018) – an endeavour pursued in
336 each of the study regions. Establishing RWLs across the study regions required addressing
337 different contextual factors and socioecological characteristics, highlighting the need to adapt
338 approaches to the unique conditions and needs of each local context. Through inclusive
339 communication and deliberation on perceptions, experiences, aspirations and expectations, a
340 shared understanding of desirable futures was developed, and these in turn guided the goals of
341 RWLs (*sensu* Leventon et al. 2016).

342 Each RWL aimed to address local issues related to grassland restoration, improve stakeholder
343 engagement, promote sustainable practices and enhance the resilience of the socioecological
344 system associated with grasslands (for a detailed discussion of the role of social factors and
345 stakeholder collaboration in enhancing restoration success, see Box 3). These objectives
346 required a nuanced understanding of spatio-temporal variations in environmental and socio-
347 economic conditions, highlighting the need for adaptive restoration strategies (Table 3). For
348 example, RWL North engaged regional stakeholders through a series of participatory
349 workshops. This approach facilitated deliberative processes and social learning that allowed
350 stakeholders to contribute to decision-making and co-create restoration strategies tailored to
351 the specific needs of the region accounting for the diversity of values and worldviews on
352 restoration. By fostering dialogue and collaboration, these workshops aimed to build trust,

353 enhance local capacity and agency for sustainable practices, and ensure that restoration efforts
354 were both contextually relevant and supported by all stakeholders. Meanwhile, the RWL Centre
355 emphasised the involvement of local stakeholders in restoration activities through citizen
356 science programmes and the monitoring of participatory pilot actions. This approach would
357 empower the community, increase scientific literacy and ensure ongoing stakeholder
358 involvement in restoration efforts. In the RWL South, activities included the creation of an
359 online forum for the community to share information about restoration projects. This digital
360 platform aimed to facilitate knowledge sharing, promote community engagement and ensure
361 transparency in restoration initiatives. Taken together, these different approaches reflect
362 adaptive, context-specific strategies in RWLs, which are critical for helping to achieve long-term
363 sustainability and resilience in grassland restoration efforts.

364 While short-term experiments within RWLs are valuable for immediate learning and adaptation,
365 they often fail to capture the complexity and long-term perspective needed to understand and
366 support sustainable transformations. Therefore, it is crucial to develop RWLs as research spaces
367 with a broad spatial, temporal and thematic scope. This broader scope enables RWLs to address
368 regional and local specificities while contributing to global knowledge, ensuring that
369 interventions are contextually relevant and widely applicable.

370 To operationalise and track advances and changes in the RWLS, we focussed on three main
371 components (see Table 3):

- 372 1. *Ex-ante/ex-post* evaluation (for the Northern and Central Regions)
- 373 2. Transdisciplinary knowledge co-creation during live restoration with local stakeholders
- 374 3. Demonstration sites

375 As a first component, the *ex-ante* evaluation measured stakeholders' initial views, including
376 their values – such as the importance they place on grasslands for ecological, cultural or
377 economic reasons – and their knowledge, referring to their understanding of grassland
378 biodiversity, ecosystem functions and restoration practices. It also assessed their motivations,
379 visions and perceived barriers related to grasslands. The *ex-post* evaluation, on the other hand,
380 involved the assessment of changes in valuation of grasslands (and nature in general) and
381 grassland restoration – also assessing the other aspects described above. The second

382 component involved transdisciplinary co-creation with local stakeholders, focusing on
383 identifying contextual issues related to live grassland restoration and co-creating knowledge
384 using co-design methods. The third component was knowledge exchange using demonstration
385 sites in all three regions to highlight multifunctional outcomes and share best practices.

386

387 **Synthesis and Integration**

388 Synthesis and integration are critical in interdisciplinary projects like *Grassworks*, where
389 combining social, ecological and economic disciplines is essential for effectively addressing
390 complex restoration challenges. By merging academic knowledge with practical expertise, we
391 aimed to create a coherent interdisciplinary framework to inform both research and practice.

392 This effort was supported by our practice partner, Deutscher Verband für Landschaftspflege
393 (DVL, Land Care Germany), who provided expert guidance and facilitated connections with
394 stakeholders across Germany (see YouTube website with films developed by the DVL to inform
395 practitioners on best practice methods:

396 <https://www.youtube.com/playlist?list=PLrA74x502hW7UKcXfjNat5zFbaSMOcgNn>). As part of
397 the synthesis, we developed a model of factors contributing to restoration success using
398 Bayesian Belief Networks (BBNs; MacPherson et al. 2018). Bayesian belief networks are acyclic
399 graphs representing networks of variables and their dependencies. The structure of our
400 Bayesian belief network was co-designed through two workshops with the *Grassworks*
401 consortium, integrating diverse perspectives with a strong focus on stakeholder views. While
402 still under development, the final BBN will enable simulations and analyses to explore how
403 changes in specific variables influence restoration success.

404 The integration of these approaches adds significant value by creating a framework that can be
405 transferred and adapted across different spatial and social contexts. The replicability of the
406 framework over time and space lies in its focus on key elements, including the consideration of
407 spatial heterogeneity in grassland systems and the inclusion of diverse stakeholder perspectives

408 across social scales. These attributes ensure the framework's applicability to different
409 restoration projects and its potential to guide long-term, sustainable restoration efforts.

410 A key outcome of *Grassworks*, as part of our synthesis and integration efforts, will be an online
411 restoration success estimation tool. Informed by ecological and social findings from
412 *Grassworks*, including the BBN analysis, this tool will provide restoration practitioners with
413 guidance and insights into the likelihood of success. Furthermore, all ecological data generated
414 during the project will be uploaded to the German GFBio biodiversity data repository
415 (<https://www.gfbio.org/materials/>) in accordance with FAIR principles (Wilkinson et al. 2016)
416 and will be made publicly available following a two-year moratorium.

417 With the EU Restoration Law and national and the EU biodiversity strategies now firmly on the
418 political agenda, the need for scaling up ecological restoration is greater than ever. Upscaling is
419 not only a question of increasing the area that is restored, but also a socialecological endeavour
420 that requires strong links and communication between science and practice as well as across
421 different social spheres of society. The *Grassworks* project has made significant progress in this
422 regard by fostering collaboration between researchers and practitioners, integrating ecological,
423 social and economic dimensions, and creating tools and frameworks designed to inform and
424 guide scalable and transferable restoration efforts. However, as with many transdisciplinary
425 projects, *Grassworks* faced its own limitations, including challenges of stakeholder engagement
426 and availability, variability of physical factors such as climate change, and administrative
427 barriers as part of bureaucratic processes. In addition, the complexity of aligning diverse
428 stakeholder interests and integrating knowledge across disciplines required considerable effort
429 and coordination. These limitations highlight the continuing need for adaptive approaches and
430 flexible frameworks to address the specific challenges of transdisciplinary and collaborative
431 research projects.

432 For restoration to be as successful as possible, attention must broaden the conventional and
433 project-based lens of ecological objectives to situating restoration as a process within a
434 socialecological system, integrating different values, practices, knowledge and goals, across
435 different stakeholder groups. While we already have substantial knowledge on the factors

436 contributing to ecological success in grassland restoration, it is important to acknowledge that
437 restoration efforts can face challenges and sometimes fail to fully achieve their goals.
438 *Grassworks* builds on this knowledge foundation by employing standardized sampling across
439 three regions and, for the first time, assessing the critical role of social as well as the holistic
440 socialecological components that drive restoration success. We consider that the outcomes
441 from the socialecological *Grassworks* project, being synthetic and integrative across a range of
442 different grassland vegetation types as well as including a broader epistemological lens, will
443 provide a strong evidence base for informing on the ground grassland restoration in Germany,
444 but also in many other countries in Europe within the framing of the EU Restoration Law. Since
445 the dynamics of grasslands across central and northern Europe are generally influenced by
446 similar drivers of degradation (intensification of land use, eutrophication, bush encroachment
447 etc.) and the need for socialecological whole system approaches to restoration are on the rise,
448 we anticipate that our findings and this methods paper should provide some key insights for
449 upcoming projects and restoration activities.

450

451 **Acknowledgements**

452 This research within the *Grassworks* project was funded by the German Federal Ministry of
453 Education and Research (BMBF) under the FEdA programme (Forschung für den Erhalt der
454 Vielfalt; <https://www.feda.bio/en/>) on Research for Biodiversity Conservation within the sub-
455 programme *BiodiWert* (grant number 16LW0095). We thank Matthias Boysen from VDI/VDE in
456 Berlin, as well as Julian Taffner and other colleagues at the FEdA headquarters for their support
457 of our endeavour. The *Grassworks* website can be found at <https://grassworksprojekt.de/en/>.
458 We are grateful to the farmers, conservation practitioners and landowners for granting access
459 to their land and to all student helpers for their invaluable support during fieldwork.
460 Additionally, we thank Greta Bindernagel for translating the survey for stakeholders in Table S2
461 and we thank Christian Schmid-Egger for helping to identify wild bees.

462

463

464 **References**

- 465 Abson DJ, J Fischer, J Leventon et al. (2017) Leverage points for sustainability transformation.
466 *Ambio* 46:30–39.
467
- 468 Bardgett , RD, JM Bullock, S Lavorel, P Manning, U Schaffnerset N Ostle, M Chomel, G Durrigan,
469 E Frey, D Johnson, J Lavallee, G Le Provost, S Luo, K Png, M Sankaran, H Xiangzang, H Zhou, L
470 Ma, R Webó, D Yong, L Yuanheng, S Hongxiao (2022) Combatting grassland degradation. *Nature*
471 *Reviews Earth &Environment*, 2 (10). 720-735.<https://doi.org/10.1038/s43017-021-00207-2>
472
- 473 Bauer M, J Huber, J Kollmann (2024) Beta diversity of restored river dike grasslands is strongly
474 influenced by uncontrolled spatio-temporal variability. *J Vegetation Science* 35:e13293.
475
- 476 Ben-Othmen A, Ostapchuk M. (2023) How diverse are farmers’ preferences for large-scale
477 grassland ecological restoration? Evidence from a discrete choice experiment. *Review of*
478 *Agricultural, Food and Environmental Studies* 104, 341–375. [https://doi.org/10.1007/s41130-](https://doi.org/10.1007/s41130-023-00200-x)
479 [023-00200-x](https://doi.org/10.1007/s41130-023-00200-x)
480
- 481 Bergmann M, N Schöpke, O Marg et al. (2021)Trandisciplinary sustainability research in real-
482 world labs: success factors and methods for change. *Sustainability Science*
483 <https://doi.org/10.1007/s11625-020-00886-8>
484
- 485 Bengtsson J, JM Bullock, B Egoh, C Everson, T Everson, T O’Connor, PJ O’Farrell, HG Smith
486 (2019) Grasslands—more important for ecosystem services than you might think.
487 *Ecosphere* Volume 10(2) e02582
488
- 489 Blickensdorfer L, M Schwieder, D Pflugmacher, C Nendel, Erasmi, P Hostert (2022) Mapping of
490 crop types and crop sequences with combined time series of Sentinel-1, Sentinel-2 and Landsat
491 8 data for Germany. *Remote Sensing of the Environment* 269 112831.
492
- 493 Broeckhoven N and A. Cliquet (2015) Gender and Restoration Ecology: time to connect the
494 dots. *Restoration Ecology* 23, 6, 729–736.
495
- 496 Brudvig LA (2017) Toward prediction in the restoration of biodiversity. *Journal of Applied*
497 *Ecology* doi: 10.1111/1365-2664.12940.
498
- 499 Bucharova A, Lampei C, Conrady M, E May, J Metheja, M Meyer, D Ott (2022) Plant provenance
500 affects pollinator network: Implications for ecological restoration. *J Appl Ecol* ;59:373–383.
501 <https://doi.org/10.1111/1365-2664.13866>
502
- 503 Buckingham K, B Arakwiye, S Ry, O Maneerattana, Anderson W (2021) Cultivating networks and
504 mapping social landscapes: How to understand restoration governance in Rwanda. *Land Use*
505 *Policy* <https://doi.org/10.1016/j.landusepol.2020.104546>.
506

507 Carlucci M, PHS Brancalion, RR Rodrigues, R Loyola, M Ciancaruso (2020) Functional traits and
508 ecosystems services in ecological restoration. *Restoration Ecology* 28: 1372-1383.
509
510
511
512 Canessa C, Venus T. E, Wiesmeier M, Mennig P, & Sauer J (2023). Incentives, rewards or both in
513 payments for ecosystem services: Drawing a link between farmers' preferences and biodiversity
514 levels. *Ecological Economics*, 213, 107954.
515
516 Choi Y., VM Temperton, EB Allen, A Grootjans, M Halassy, RJ Hobbs, MA NAeth and K Torok
517 (2008) Ecological restoration for future sustainability. *Ecoscience* 15: 53-64.
518
519 Chytrý M., Tichý, L., Hennekens, S. M., Knollová, I., Janssen, J. A. M., Rodwell, J. S., Peterka, T.,
520 Marcenò, C., Landucci, F., Danihelka, J., Hájek, M., Dengler, J., Novák, P., Zukal, D., Jiménez-
521 Alfaro, B., Mucina, L., Abdulhak, S., Ačić, S., Agrillo, E., ... Schaminée, J. H. J. (2020). EUNIS
522 Habitat Classification: Expert system, characteristic species combinations and distribution maps
523 of European habitats. *Applied Vegetation Science*, 23(4), 648–675;
524 <https://doi.org/10.1111/avsc.12519>
525
526
527 Drachenfels OV (2021) Kartierschlüssel für Biotoptypen in Niedersachsen unter
528 Berücksichtigung der gesetzlich geschützten Biotope sowie der Lebensraumtypen von Anhang I
529 der FFH-Richtlinie. *Naturschutz und Landschaftspflege in Niedersachsen*, Heft A/4,
530 Niedersächsischer Landesbetrieb für Wasserwirtschaft, Küsten-und Naturschutz.
531
532 De Groot RS, J Blignault, S van der Ploeg, J Aronson, T Elmqvist, J Farley (2013) Benefits of
533 Investing in Ecosystem Restoration. *Conservation Biology* 27, No. 6, 1286–1293.
534
535 Dengler J and S Tischew (2018) Grasslands of western and northern Europe – between
536 intensification and abandonment. In: *Grasslands of the World* (edited by Squires, Dengler, Fend
537 and Hua) International Standard Book Number 13: 13: 1-4987-9626-2
538
539 Dengler J, Janisova M, Török P, Wellstein C (2014) Biodiversity of Palaearctic grasslands: a
540 synthesis. *Agriculture Ecosystems and Environment* 182: 1-14.
541
542 Dullau S, A Kirmer, Sabine Tischew, F Holz, MH Meyer, A Schmidt (2023)
543 Effects of fertilizer levels and drought conditions on species assembly and biomass production
544 in the restoration of a mesic temperate grassland on ex-arable land. *Global Ecology and*
545 *Conservation* 48: e02730
546
547 Elias M, D Joshi, & R Meinzen-Dick (2021) Restoration for whom, by whom? A feminist political
548 ecology of restoration. *Ecological Restoration* 39: 3-15.
549

550 Fernández-Manjarrés JF, S Roturier, & AG Bilhaut (2018) The emergence of the social-ecological
551 restoration concept. *Restoration Ecology*, 26(3), 404-410.
552

553 Fischer J, M Riechers, J Loos, B Martin-Lopez, VM Temperton (2021). Making the UN Decade on
554 Ecosystem Restoration a Social-Ecological Endeavour. *Trends in Ecology & Evolution* 36(1), 20-
555 28, DOI: 10.1016/j.tree.2020.08.018.
556

557 Funk JL, Larson JE, Ames GM, Butterfield BJ, Cavender-Bares J, Firn J, Laughlin DC, Sutton-Grier
558 AE, Williams L, Wright J (2017) Revisiting the holy grail: using plant functional traits to
559 understand ecological processes. *Biological Reviews* 92:1156–
560 1173. <https://doi.org/10.1111/brv.12275>
561

562 Hellwig, N., Schubert, L. F., Kirmer, A., Tischew, S., & Dieker, P. (2022). Effects of wildflower
563 strips, landscape structure and agricultural practices on wild bee assemblages—A matter of data
564 resolution and spatial scale? *Agriculture, Ecosystems & Environment*, 326, 107764.
565

566 Helm A, M Zobel, AT Moles, R Szava-Kovats & M Partel (2015)
567 Characteristic and derived diversity: implementing the species pool concept
568 to quantify conservation condition of habitats. *Diversity and Distributions* 21:711-721.
569

570 Hess MCM, E. Buisson, H. Fontes, L Bacon, F Sabatier, F Mesléard (2020) Giving recipient
571 communities a greater head start and including productive species boosts early resistance to
572 invasion. *Applied Vegetation Science*
573

574 Huber, R., & Finger, R(2020). A meta-analysis of the willingness to pay for cultural services from
575 grasslands in Europe. *Journal of Agricultural Economics*, 71(2), 357-383.
576

577 Jandt U, Bruelheide H, Jansen F, Bonn A, Grescho V, Klenke RA, Sabatini FM,
578 Bernhardt-Römermann M, Blüml V, Dengler J (2022) More losses than
579 gains during one century of plant biodiversity change in Germany. *Nature*
580 611:512–518. <https://doi.org/10.1038/s41586-022-05320-w>
581

582 Knight M. L., Overbeck G. E. (2021) How much does it cost to restore a grassland? *Restoration*
583 *Ecology* 29 (8), e13463. <https://doi.org/10.1111/rec.13463>
584

585 Leuschner C, K Wesche, S Meyer, B Krause, K Steffen, T Becker, & H Culmsee
586 (2013). Veränderungen und Verarmung in der Offenlandvegetation Norddeutschlands
587 seit den 1950er Jahren: Wiederholungsaufnahmen in Äckern, Grünland und
588 Fließgewässern. In *Bericht. d. Reinard Tüxen Gesellschaft* (Vol. 25).
589

590 Leventon J; L Fleskens, H Claringbould, G Schwilch and R Hessel (2016) An applied methodology
591 for stakeholder identification in transdisciplinary research. *Sustainability Science* 11:763-775.
592

593 Lyons, KG, P Török, J-M Hermann, K Kiehl, A Kirmer, J Kollmann, GE Overbeck, S Tischew, EB
594 Allen, JD Bakker, C Brigham, E Buisson, K Crawford, P Dunwiddie, J Firn, D Grobert, K Hickman, S
595 Le Stradic, VM Temperton (2023). Challenges and opportunities for grassland restoration: A
596 global perspective of best practices in the era of climate change. *Global Ecology and*
597 *Conservation*, 46, e02612, DOI: 10.1016/j.gecco.2023.e02612
598

599 Martin DM (2017). Ecological restoration should be redefined for the twenty-first century.
600 *Restoration Ecology. Vol. 25, No. 5, pp. 668–673*
601

602 MacPherson MP, EB Webb, A Raedeke, D Mengel, F Nelson (2018) A review of Bayesian belief
603 network models as decision-support tools for wetland conservation: are water birds potential
604 umbrella taxa? *Biological Conservation* 226
605

606 Messier C, K Puettman, R Chazdon, KP Andersson, VA Angers, L Brotons, E Filotas, R Tittler, L
607 Parrott, SA Levin (2014) From Management to Stewardship: Viewing Forests as Complex Adaptive
608 Systems in an Uncertain World. *Conservation Letters* 8:368-377
609

610 Montoya D, L Rogers and J Memmott (2012) Emerging perspectives in the restoration of
611 biodiversity-based ecosystem services. *Trends in Ecology and Evolution*.
612

613 Nerlekar AN, Sullivan LL, Brudvig LA (2024) Grassland restorations must better foster forbs to
614 facilitate high biodiversity. *Restoration Ecology* doi: 10.1111/rec.14214
615

616 NLWKN (Hrsg.) (2011): Vollzugshinweise zum Schutz der FFH-Lebensraumtypen sowie weiterer
617 Biototypen mit landesweiter Bedeutung in Niedersachsen. – FFH-Lebensraumtypen und
618 Biototypen mit höchster Priorität für Erhaltungs- und Entwicklungsmaßnahmen –
619 Niedersächsische Strategie zum Arten- und Biotopschutz, Hannover

620 Pfadenhauer J (2001) Some Remarks on the Socio-Cultural Background of Restoration Ecology.
621 *Restoration Ecology* Vol. 9 No. 2, pp. 220–229
622

623 Pfadenhauer J, Poschlod P & Buchwald R (1986). Überlegungen zu einem Konzept
624 geobotanischer Dauerbeobachtungsflächen für Bayern. Teil I. *Berichte der Bayerischen*
625 *Akademie für Naturschutz und Landschaftspflege*, 10, 41–60. Retrieved from
626 [https://www.zobodat.at/pdf/Ber-Bayer-Akad-f-Natursch-u-Landschaftspfl_10_1986_0041-](https://www.zobodat.at/pdf/Ber-Bayer-Akad-f-Natursch-u-Landschaftspfl_10_1986_0041-0060.pdf)
627 [0060.pdf](https://www.zobodat.at/pdf/Ber-Bayer-Akad-f-Natursch-u-Landschaftspfl_10_1986_0041-0060.pdf)
628

629 Pfadenhauer J (2001) Some Remarks on the Socio-Cultural Background of Restoration Ecology.
630 *Restoration Ecology* 9: 220-229
631

632 Perring MP, TE Erickson, P Brancalion (2018) Rocketing restoration: enabling the upscaling of
633 ecological restoration in the Anthropocene. *Restoration Ecology* doi: 10.1111/rec.12871
634 Shackelford N, RJ Hobbs, JM Bruger et al. (2013) Primed for Change: Developing Ecological
635 Restoration for the 21st Century. *Restoration Ecology* Vol. 21, No. 3, pp. 297–304.
636

637 Shackelford N, RJ Hobbs, JM Burgar, TE Erickson, JB Fontaine, E Lalliberte, CE Ramalho, MP
638 Perring, RJ Standish (2013) Primed for Change: Developing Ecological Restoration for the 21st
639 Century. *Restoration Ecology* 21: 297-304.
640
641 Shackelford N BB Paterno, DE Winkler et al. (2021a) Drivers of seedling establishment success in
642 dryland restoration efforts. *Nature Ecology and Evolution*. VOL 5: 1283–1290
643
644 Shackelford N, Dudney J, Stueber MM, Temperton VM, Suding KL (2021b) Measuring at all
645 scales: sourcing data for more flexible restoration references. *Restoration Ecology* doi:
646 10.1111/rec.13541
647
648 Socher SA, Prati D, Boch S, Müller J, Baumbach H, Gockel S, Hemp A, Schöning I, Wells K, Buscot
649 F, Kalko EKV, Linsenmair KE, Schulze E-D, Weisser WW, Fischer M (2013) Interacting effects of
650 fertilization, mowing and grazing on plant species diversity of 1500 grasslands in Germany differ
651 between regions. *Basic and Applied Ecology* 14(2) 126-136.
652
653 Schöpke N, F Stelzer, G Caniglia et al. (2018) Jointly experimenting for transformation?
654 Shaping real-world laboratories by comparing them. *GAIA-Ecological Perspectives for Science
655 and Society*, 27(1), 85-96.
656
657 Schöpke N, R Beecroft, M Wanner, F Wagner, R Rhodius, P Laborgne and O Parodi. (2021)
658 Gaining deep leverage? Reflecting and shaping real-world lab impacts through leverage points.
659 *Gaia* 33:116-124.
660
661 Shipley JR, ER Frei, A Bergmaini, N Pichon, et al. (2019) Agricultural practices and biodiversity:
662 Conservation policies for semi-natural grasslands in Europe. *Current Biology* 34: 747-771.
663
664 Schwieder, M Wesemeyer, D Frantz, K Pfoch, S Ersami, J Pickert, C Nendel, P Hostert (2022)
665 Mapping grassland mowing events across Germany based on combined Sentinel-2 and Landsat
666 8 time series. *Remote Sensing of the Environment* 269: 112795.
667
668 Squires VR, J Dengler, H Feng, L Hua (2018) *Grasslands of the World: Diversity Management and
669 Conservation*. CRC Press, 42 pages, International Standard Book Number-13: 1-4987-9626-2
670
671 Staude IR, J Segar VM Temperton, BO Andrade, M de Sá Dechoum, EWA Weidlich, GE Overbeck
672 (2023). Prioritize grassland restoration to bend the curve of biodiversity loss. *Restoration
673 Ecology*, 31 (5), e13931, DOI: 10.1111/rec.13931.
674
675 Stelzenmüller V; HO Fock, A Gimpel, H Rambo, R Diekmann, WN Porbst, U Callies, F
676 Bockelmann, H Neumann, I Kröncke (2015) Quantitative environmental risk assessments in the
677 context of marine spatial management: current approaches and some perspectives. *ICES
678 Journal of Marine Science* 72(3), 1022–1042.
679

680 Tedesco AM, S López-Cubillos, R Chazdon et al. (2023) Beyond ecology: ecosystem restoration as
681 a process for social-ecological transformation. Trends in Ecology and Evolution Vol. 38, No. 7.
682
683 Tischew S, A Baasch, MK Conrad, A Kirmer (2008) Evaluating Restoration Success of Frequently
684 Implemented Compensation Measures: Results and Demands for Control Procedures.
685 Restoration Ecology
686
687 Tivian. (2024). Online survey software: Surveys made easy with Unipark Tivian XI GmbH.
688 <https://www.unipark.com/en/>
689
690 Traveset A, Lara-Romero C, Santamaria S, Escribano-Avila G, Bullock JM, Honnay O, Hooftman
691 DAP, Kimberly A, Krickl P, Plue J, Poschlod P, Cousins SAO (2023) Effect of green infrastructure
692 on restoration of pollination networks and plant performance in semi-natural dry grasslands
693 across Europe. J. Applied Ecology. DOI: 10.1111/1365-2664.14592
694
695 Wanner M, A Hilger, J Westerkowski, M Rose, F Stelzer & N Schöpke (2018) Towards a Cyclical
696 Concept of Real-World Laboratories: A Transdisciplinary Research Practice for Sustainability
697 Transitions. disP –The Planning Review 54(2): 94-114.
698
699 Waldén E & R Lindborg (2018) Facing the future for grassland restoration – What about the
700 farmers? Journal of Environmental Management 227: 305-312.
701
702 Wesche K, Kruase B, Culmsee H, Leuschner C (2012) Fifty years of change in Central European
703 grassland vegetation: Large losses in species richness and animal-pollinated plants. Biological
704 Conservation 150:76-85.
705
706 Wilkinson (2016) Comment: The FAIR Guiding Principles for scientific data management and
707 stewardship- Nature Scientific Data 3:160018 | DOI: 10.1038/sdata.2016.18
708
709 Wortley L, JM Hero & M Howes (2013) Evaluating Ecological Restoration Success: A Review of
710 the Literature. Restoration Ecology Vol. 21, No. 5, pp. 537–543
711
712 Zedler J (2007) Success: An Unclear, Subjective Descriptor of Restoration Outcomes. Ecological
713 Restoration 25 (3).
714
715 Zerbe S (2023) Restoration Economy: Costs and benefits. In: Restoration of Ecosystems-
716 Bridging Nature and Humans. Springer Spektrum, Berlin, Heidelberg.

717

718 **Tables**

719

720 **Table 1.** Overview of site and management variables used as stratification factors for the landscape-scale post-hoc analysis of
721 restoration success. These site and management variables and their levels were developed using expert knowledge within the
722 consortium as well as a survey distributed to restoration stakeholders across the three regions (51 people filled out the survey about
723 a total of 183 sites, including reference sites). Wherever possible we tried to have similar numbers of sites per level within a region.

724

725 Variable	726 Level			
727 <i>Site</i>				
727 Aim of restoration	species-rich	erosion control	carbon storage	landscape connectivity
728 Hydrology	dry	fresh	moist	
729 Previous land use	grassland	arable land		
730 Time since restoration	1–5 years	6–10 years	>10 years	
731 <i>Management</i>				
732 Site preparation	creation of open soil	nutrient reduction	shrub removal	
733 Restoration measure	cultivar seed mix	regional seed mix	direct harvesting	management adaptation
734 Current management	grazing	mowing	grazing and mowing	

735

736 **Table 2.** Number of sites sampled across the three regions in the post-hoc assessment in *Grassworks*, showing the number of
 737 restored, positive and negative reference sites sampled per region (187 sites in total) in the first section of the table. The second
 738 section shows values for the restored sites (without references sites) for each level of each stratification factor (restoration method,
 739 previous land use, and age since restoration) giving a total of 122. It took two growing seasons to measure all sites (2022 and 2023).
 740 *Restoration method* abbreviations: MgA = management adaptation; DiH = direct harvesting, ReS = regional seed mixture, CuS =
 741 cultivar seed mixture.

Variable	<i>North</i>	<i>Centre</i>	<i>South</i>	<i>Subtotal</i>
<i>Site type</i>				
restored	40	41	40	121
positive	11	12	10	33
negative	10	13	10	33
<i>Current Management</i>				
mowing	35	32	57	124
grazing	12	17	1	30
both	13	10	2	25
no	1	7	0	8
<i>Subtotal (all sites)</i>	61	66	60	187

<i>Restoration method</i>				
MgA	8	14	0	22
DiH	11	4	25	40
ReS	10	17	11	38
CuS	11	6	4	21
<i>Previous land use</i>				
grassland	13	15	13	41
arable land	27	26	27	80

Age since restoration

(years)

<i><5</i>	<i>18</i>	<i>12</i>	<i>9</i>	<i>39</i>
<i>6-10</i>	<i>11</i>	<i>11</i>	<i>6</i>	<i>28</i>
<i>>10</i>	<i>9</i>	<i>15</i>	<i>17</i>	<i>41</i>
<i>NA</i>	<i>2</i>	<i>3</i>	<i>8</i>	<i>13</i>

Subtotal (restored sites) *40* *41* *40* *121*

742

743

744

745

746

747

748 **Table 3.** Key differences in characteristics and methodological approaches of Real World Laboratories (RWLs) in the *Grassworks*
 749 project

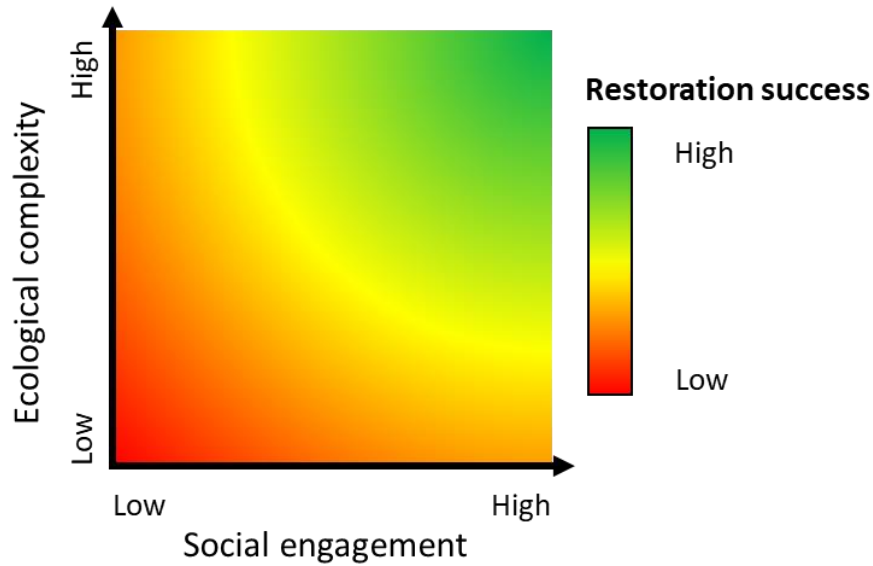
Real-World Laboratory (RWL) Characteristics	RWL North	RWL Centre	RWL South
Case study region	Landkreis Gifhorn, focus on Hankensbüttel and Drömling	Biosphere Reserve Karst Landscape Southern Harz, focus on Hainrode village	Donau-Isar lowland, Southern Bavaria, focus on Gauting
Cooperation partner	Gifhorn County, Aktion Fischotterschutz e., and Biohof Flegel KG	Heimat and Naturschutzverein Hainrode e. V.	Gauting Municipality
Economic characteristics of the region	Medium economic strength	Low economic strength	High economic strength
Biophysical characteristics of the region	Medium nutrient inputs and extensive grasslands owned by the cooperation partner are species-poor and degraded despite restoration efforts. Species-rich grasslands are scarce in the model region.	Increasing risk of grassland fallow and loss of connectivity. However, target species and residual populations remain, indicating good regeneration potential with low nutrient inputs.	Intensive agricultural use, high nutrient inputs, fertile soils. Grasslands threatened by ploughing, gravel extraction, and dyke construction.
Demonstration site(s)	Long-term grassland experiment testing potential of priority effects (POEM project, Alonso-Crespo et al. in revision) for multifunctional outcomes.	Long-term species-rich grassland fertilization experiments in Hayn (Saxony-Anhalt; Dullau et al. 2023) aiming at high plant biodiversity and productivity outcomes.	Long-term dike restoration experiment (Bauer et al. 2024) testing site characteristics, spatial and historical effects on restoration outcomes near the Danube (Lower Bavaria)
Live Restoration	<ul style="list-style-type: none"> a) Rewetting of grasslands, and testing potential for using biodiversity and priority effects for multifunctional outcomes (Hankensbüttel). b) Restoration intervention with hay transfer from donor to receiver sites as part of compensation for building (Drömling) 	Establishment of five species-rich grassland areas in Hainrode, accompanied by informational and educational materials (e.g., informational signs, a citizen science brochure, distribution of seed packets, and an information booth at village events).	Improvement of a municipal grassland site through a simple experiment: sowing regional seed at 2 densities and with 4 soil preparation methods (topsoil removal, grubbing, rotary tilling, management adaption/no preparation); Included a student project and resident's engagement through an online forum, informational signs, and a live event (Feldtag).
Approaches and methods used in the RWL	Transdisciplinary process for the co-creation and implementation of a long-term pilot restoration concept in collaboration with local stakeholders. Value-based envisioning workshop, restoration concept world café workshop, ex-ante/ex-post comparison, qualitative interviews, group deliberative Q-method, and social network analysis.	Participatory mapping exercise for the site selection process (workshop), vegetation surveys, photovoice, focus groups, citizen science activities, and ex-ante/ex-post surveys.	Vegetation surveys before/after sowing, photo documentation; Discussions through online forum, ex-ante/(ex-post is pending); Field day

750

751

752

753 **Figures**

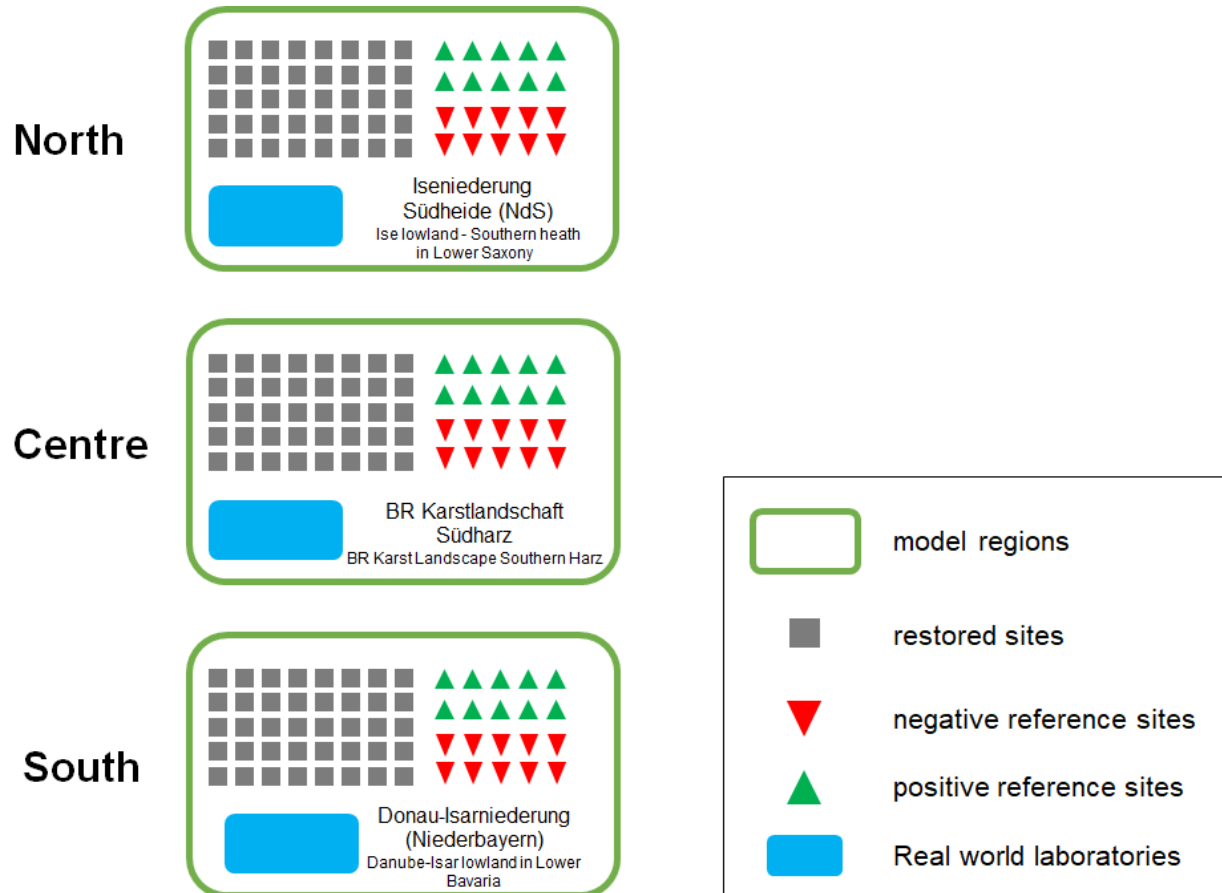


754

755

756 Figure 1. Our central hypothesis is that the overall sociaecological success of restoration relates to the extent to which both
757 ecological and stakeholder complexity are considered in the restoration process. We hypothesise that the higher the ecological
758 complexity and social engagement (stakeholder diversity/inclusion) are, the higher the restoration success will be. Our combination
759 of ecological, sociaecological and economic data within *Grassworks* allows us to test this hypothesis. The success of restoration is

760 shown as a sequence of colours, ranging from low (red) to middle (yellow) and high levels (green), the latter can only be achieved
761 with high values on both social and ecological axes.



762

763

764 Figure 2. The *Grassworks* research approach for assessing grassland restoration success using a natural landscape experiment
765 approach for the post-hoc sampling and real-world laboratories (RWL). We compared three regions from north to south in Germany.
766 Within each region we did a post-hoc assessment of approximately 40 already restored grasslands as well as ten positive and ten

767 negative reference sites (see Table 2 for exact details of numbers of sites per region and category). In addition, we set up one RWL
768 per region, where *in situ* restoration with local stakeholders was co-designed and performed (for the North and the Centre) and an
769 online forum was established in the South. BR = biosphere reserve.

770



771

772

773 Figure 3. Approach to assess ecological parameters at each of the restored or reference sites in the post-hoc assessment in three
 774 regions (Figure 2). When the grassland site was long enough, we worked along a 200 m transect with vegetation plots (4 m²) every
 775 50 m. At sites with different spatial formats, we sampled across four separate 50 m transects (not shown). The dark green quadrats
 776 denote areas where vegetation cover was assessed in detail; lighter green areas denote where a full plant species richness was
 777 collected as well as where butterflies and wild bees were sampled. Red dots show the locations of pooled soil samples taken for
 778 total carbon (SOC) and nitrogen, pH, soil texture, and microbial biomass (carbon-based). Insects were sampled four times over a
 779 growing season compared to vegetation once.

780

781

782 **Boxes**

783 Box 1. Reflections on selecting restored sites and what constitutes a restoration method within the post-hoc approach

784 When selecting sites in the three regions as part of our natural landscape experiment approach (Figure 2) we decided to take the
785 availability of different grassland types in the real landscape within our three larger regions (North, Centre, South) into account, and
786 did not stringently balance every factor that we considered important for restoration outcomes (see Table 1 as well as Table S2 for
787 more details). This particularly applied to some regionally more specific factors that are affected by local climate (e.g. prevalence of
788 wetter grasslands in the North compared to Central or Southern Germany), cultural and historical land use (e.g. less grazing in the
789 southern region than in the North or the Centre, with sites in the South concentrated around the lowlands surrounding Munich in
790 Lower and Upper Bavaria.

791 Nevertheless, when selecting sites, we strove to balance out the number of sites per level as much as possible for the following main
792 factors: hydrology, time since restoration intervention, restoration method, previous land use and current management (Table 1,
793 Table S2). A certain number of factors were deemed potentially important but too difficult to adequately assess prior to starting
794 fieldwork. Here the information was obtained later and included the full variety of funding instruments, whether a grassland site
795 was located in a nature conservation area or not, different soil preparation approaches before restoration. We found restored sites
796 via our networks of local contacts, previous collaboration partners in conservation practice, local conservation authorities and NGOs
797 and through a snowballing effect amongst these stakeholders.

798 The question of whether management (grazing or mowing) constitutes a restoration method (as opposed to direct harvesting or
799 sowing or seeds) engendered a lively debate. Whilst all grasslands that are not on extremely wet or dry sites, require some form of
800 disturbance (grazing or mowing) to remain grasslands and not go through successional processes, one form of ecological restoration

801 of grasslands includes adapting such grazing or mowing management. As such, we decided to include a combination of mowing and
802 grazing as one of the restoration methods in our post-hoc analysis of the restored and reference sites.

803

804 **Box 2.** Reflection on selecting positive and negative reference sites

805 Ecological restoration often compares the outcomes of a restoration intervention to a contemporary positive reference site. The
806 contemporary positive reference generally represents the desired ecosystem state, usually with respect to plant species composition
807 and diversity, vegetation structure and ecosystem functions, as well as sometimes forb to grass ratio in grassland restoration. The
808 extent to which a restored site is converging on the plant species compositional space of a positive reference (or not) in multivariate
809 analyses conventionally forms part of the method to monitor the success of restoration projects (e.g. Choi et al. 2008). Comparing
810 restoration outcomes to negative (degraded) reference sites is much less commonly done (Shackelford et al. 2021b) but can frame
811 the overall trajectory comparisons effectively (Wortley et al. 2013). Drivers of degradation in species-rich grasslands are intensive
812 land use (such as fertilizing, or mowing more than twice a year), or abandonment and subsequent shrub and tree encroachment
813 (Shiple et al. 2019). For an overview of the use of contemporary, historical or future reference sites across local to regional scales
814 see Shackelford et al. (2021).

815 In *Grassworks*, we used contemporary and local reference sites, both negative and positive. We chose classical contemporary
816 positive reference sites based on their vegetation composition and diversity since these were more easily available and since data
817 on other attributes of the reference sites (e.g. functions, functional traits, ecosystem functions and services) were too sparsely
818 available (see Funk et al. 2023 on this topic). One of our main aims was to compare the restoration outcomes with positive and
819 negative reference sites and thus assess how the inclusion of negative references affects the visualisation of restoration success. To
820 categorize what constitutes a positive reference site we used both EU and German state- level information on vegetation of
821 different grassland habitat types including grasslands within the EU Flora Fauna Habitats Directive and Natura 2000, but also
822 regional environment ministry databases (e.g. NLWKN 2021; [https://www.nlwkn.niedersachsen.de/vollzugshinweise-arten-
823 lebensraumtypen/vollzugshinweise-fuer-arten-und-lebensraumtypen-46103.html](https://www.nlwkn.niedersachsen.de/vollzugshinweise-arten-lebensraumtypen/vollzugshinweise-fuer-arten-und-lebensraumtypen-46103.html)). In addition, we used the following regional sources of

824 habitat information (Drachenfels 2021, regional lists of donor sites (Spenderflächenkataster:
825 <https://www.spenderflaechenkataster.de/startseite>), Landesamt für Umweltschutz (LAU) in Saxony-Anhalt (2019; FFH-
826 Lebensraumtypen in Appendix I of Fauna-Flora-Habitat-Directive (Directive 92/43/EWG; [https://www.lvermgeo.sachsen-
827 anhalt.de/de/gdp-geodaten-karten.html](https://www.lvermgeo.sachsen-
827 anhalt.de/de/gdp-geodaten-karten.html)), and Biotopkartierung Bayern
828 (<https://www.lfu.bayern.de/natur/biotopkartierung/index.htm>).

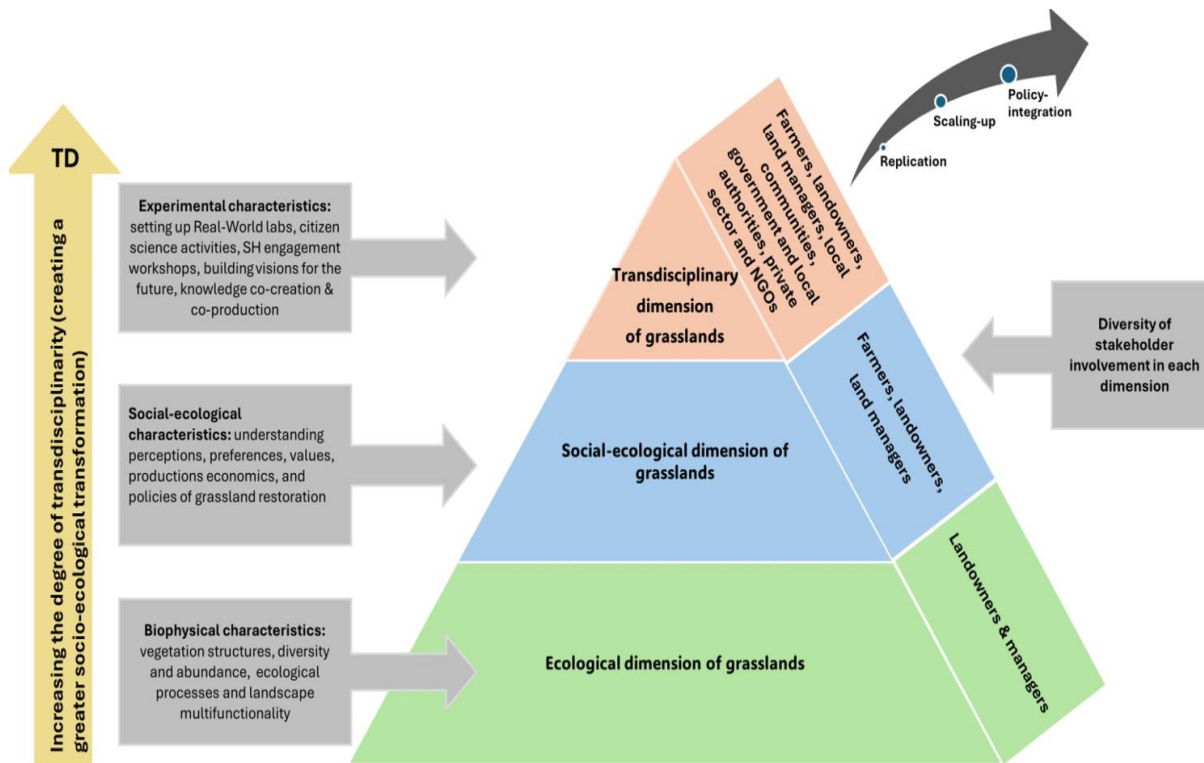
829 Our approach has the strength that our restored and reference sites represent a gradient of restoration intensity across three
830 different regions in Germany. The challenge however, consisted of finding both positive and negative reference sites. Initially, we
831 had hoped to be able to pair the restored sites with one negative and one positive reference site each, but it was not possible to
832 obtain enough negative or positive reference sites from our network of stakeholders. Overall, finding restored sites proved easier
833 than finding positive reference sites, with negative reference sites being the hardest to find. Presumably, good quality positive
834 reference sites are now rather rare, and for the negative references, it seems that many stakeholders and organizations were
835 reluctant to provide us with degraded, low diversity grassland sites, which was an interesting realisation during the process of site
836 selection during the planning of the post-hoc site analysis.

837 **Box 3.** How social dimensions of restoration complement studying restoration success, by including social perspectives and social
838 factors that influence processes and outcomes.

839 As one moves from ecological to socialecological and finally to the transdisciplinary dimensions of grassland restoration, the
840 diversity of stakeholders and the potential for socialecological transformation increase. The transdisciplinary dimension integrates
841 experimental approaches such as real-world laboratories (RWL), citizen science activities, stakeholder workshops, and knowledge
842 co-creation, fostering collaboration, trust and shared understanding among diverse stakeholders, including farmers, landowners,
843 local communities, policymakers, and NGOs.

844 This integration of ecological, social-ecological, and transdisciplinary elements enhances the capacity for systemic and lasting change
845 in grassland restoration. By addressing diverse values, practices, and knowledge systems, and through co-creation and shared
846 decision-making, the process supports scaling up, identifying best practices, and embedding key findings into policy frameworks.

847 We view this gradient, culminating in transdisciplinary approaches, as a pathway toward stronger transformation with increased and
848 more persistent restoration success. Such approaches are critical to addressing the current need for scaling up restoration efforts, as
849 framed by the EU Restoration Law and the Kunming-Montreal Biodiversity Agreement.



A holistic perspective on social-ecological system dimensions within the Grassworks project

850
851
852
853
854
855
856
857

Socioecological approaches have clear advantages over the more traditional method of assessing restoration success based on ecological attributes of the ecosystem alone, in that factors that may be critical to the chances of a project being successful may lie as much in the realm of social components (framings, values held, priorities of stakeholders, capacity for monitoring before and after effects, power dynamics, network interactions) as in the level of ecological or biophysical drivers considered or ecological attributes measured.

858 Transdisciplinary socialecological approaches take time however, as they, by definition, include more participants, who come from
859 different backgrounds and may have different knowledge or value bases (Schäpke et al. 2021). There are indeed multiple levels of
860 interdisciplinarity both within more natural science focused research projects but also within socialecological research framings as
861 *Grassworks*, where not only the language used by scientists but the methods and approaches to extracting knowledge can differ
862 ostensibly. Having a large number of social scientists working in the same project, often within the same work-packages, allows
863 strong standardisation potential but also potential for conflicting needs in relation to access to stakeholder for interviewing or
864 surveying. This requires a high level of openness and exchange, as well as time.

865

866

867

868

869

870

871

872

873

874

875

876 **Supporting Information**

877 Table S1: Braun Blanquet Plant Cover Scale. We used the following plant cover scale in *Grassworks*, adapted from the original Braun-
878 Blanquet scale, for vegetation surveys on 4 m² quadrats. This scale follows (Pfadenhauer et al., 1986) but has some adapted cover
879 and mean coverages.

Cover class	Cover [%]	Mean cover [%]
r	< 0.1	0.1
+	0.1 - 1	0.5
1a	1- 3	2
1b	3 -5	4
2a	5 -15	10
2b	15 - 25	20
3	25 - 50	37.5
4	50 - 75	62.5
5	75 - 100	87.5

880

881

882

883

884 Table S2: Overview of questions asked to stakeholders in preparation for identifying key factors affecting restoration success
885 outcomes as well as key components of restoration projects.

886

<i>Category</i>	<i>Questions / Possible answers</i>
(A) Contact	Name

	<p>Name of the farm or organisation</p> <p>Address</p> <p>Telephone number</p> <p>E-mail address</p> <p>Would you like us to contact you for future research in the field of grassland restoration?</p> <p>How large is the area that has been restored?</p> <p>What is the water balance of the area?</p> <p>It is very relevant for us to assess the location of the restored area in the landscape. We do this using aerial photographs or satellite images. Please provide information on the location of the area. You are welcome to copy the coordinates from google maps or other software here.</p>
(B) Owner	Who is the owner of the restoration area (or reference site)?
(C) Objectives	<p>Establishment of species-rich grassland</p> <p>Habitat network</p> <p>Erosion control</p> <p>Carbon sequestration</p> <p>Further objective 1</p> <p>Further objective 2</p>
(D) Success of the objectives	<p>Establishment of species-rich grassland</p> <p>Habitat network</p> <p>Erosion control</p> <p>Carbon sequestration</p> <p>Further objective 1</p> <p>Further objective 2</p>
(E) Utilization before restoration	<p>How was the area used before the restoration?</p> <p>How long has it been since the area was restored?</p>

(F) Preparation prior to restoration	<p>Scrub clearance: Have woody plants been removed?</p> <p>Soil Nutrient Depletion: Have nutrients been removed from the nutrient-rich (arable) soil, e.g. by growing crops without fertiliser? Creation of open soil: Has the area been tilled, e.g. by tilling/grubbing, rotovating or ploughing?</p>
(G) Restoration method	<p>Wild plants from certified propagation (regional seed mixture; Regiosaatgut in German)</p> <p>Direct harvesting methods such as green hay transfer</p> <p>Cultivar seed mixtures (Regelsaatgut in German)</p> <p>Other</p> <p>Is there an area nearby where the grassland develops without sowing and without fertilisation?</p> <p>How is the area managed or maintained?</p> <p>Does monitoring take place to document changes in the area?</p>
(H) Funding set-up	<p>How or under which programme were the area's restoration measures funded?</p> <p>Other</p>
(I) Funding for management	<p>How is the ongoing maintenance of the area and follow-up management funded?</p> <p>Other</p>
(J) Control	<p>Who has checked whether the planned restoration measures are being implemented correctly?</p>

888 Table S3: Overview of data sources for retrieving remote sensing data on surrounding landscape around restored grassland sites.
 889 Strg + click to open links

Federate State	Used Data	Data-sources
Schleswig-Holstein	Official Real Estate Cadastre Information System (ALKIS)	ALKIS: Landesamt für Vermessung und Geoinformation Schleswig-Holstein (https://geodaten.schleswig-holstein.de/gaialight-sh/apps/dl/download/dl-alkis.html) Biotope Maps: Landesamt für Landwirtschaft, Umwelt und ländliche Räume des Landes Schleswig-Holstein (https://opendata.schleswig-holstein.de/dataset/biotopkartierung).
Lower Saxony	German official Digital Landscape Model Base (Basic DLM) + Integrated Administration and Control System (IACS) + OpenStreetMap	Basic DLM retrieved via Thünen Institute for Biodiversity, Braunschweig. IACS: LEA Portal (https://sla.niedersachsen.de/landentwicklung/LEA/). OSM Data via R-Package „osmdata“.
Hamburg	Official Real Estate Cadastre Information System (ALKIS)	ALKIS: Landesbetrieb Geoinformation und Vermessung Hamburg (https://metaver.de/trefferanzeige?docuuid=DC71F8A1-7A8C-488C-AC99-23776FA7775E). Biotope Maps: https://suche.transparenz.hamburg.de/dataset/biotopkataster-hamburg10?forceWeb=true
Saxony-Anhalt	Official Real Estate Cadastre Information System (ALKIS)	ALKIS: Landesamt für Vermessung und Geoinformation Sachsen-Anhalt (LVerGeo) (via personalized Download). Biotope Mapping Data: Landesamt für Umweltschutz Sachsen-Anhalt (LAU) (via personalized Download).

Bavaria	Official Real Estate Cadastre Information System (ALKIS) Biotope Mappings.	ALKIS: Bayerische Vermessungsverwaltung (https://geodatenonline.bayern.de/geodatenonline/) Biotope Maps: Bayerisches Landesamt für Umwelt (https://www.lfu.bayern.de/natur/biotopflaechen_sachdaten/index.htm)
Brandenburg	Official Real Estate Cadastre Information System (ALKIS)	ALKIS: Landesvermessung und Geobasisinformation Brandenburg (https://data.geobasis-bb.de/geobasis/daten/alkis/Vektordaten/shape/)
Mecklenburg-West Pomerania	Official Real Estate Cadastre Information System (ALKIS) Biotope Mappings.	ALKIS: Landesamt für innere Verwaltung - Amt für Geoinformation, Vermessungs- und Katasterwesen Mecklenburg-Vorpommern (via personalized download). Biotope Maps: Landesamt für Umwelt, Naturschutz und Geologie Mecklenburg Vorpommern (via CD).

891

892